

**Journal of Environmental & Earth Sciences**

https://journals.bilpubgroup.com/index.php/jees

### **ARTICLE**

# **Enhanced Lead and Zinc Removal via Prosopis Cineraria Leaves Powder: A Study on Isotherms and RSM Optimization**

*Rakesh Namdeti 1[\\*](https://orcid.org/0000-0001-5255-6834) , Gaddala Babu Rao <sup>2</sup> [,](https://orcid.org/0000-0003-4190-8325) Nageswara Rao Lakkimsetty <sup>3</sup> [,](https://orcid.org/0000-0002-5042-515X) Noor Mohammed Said Qahoor <sup>1</sup> , Naveen Prasad B. S [1](https://orcid.org/0000-0001-8720-445X) , Uma Reddy Meka <sup>1</sup> [,](https://orcid.org/0009-0009-9924-2168) Prema. P.M <sup>1</sup> [,](https://orcid.org/0009-0005-9531-7717) Doaa Salim Musallam Samhan Al-Kathiri <sup>1</sup> , Muayad Abdullah Ahmed Qatan <sup>4</sup> , Hafidh Ahmed Salim Ba Alawi <sup>1</sup>*

*<sup>1</sup> Chemical Engineering, College of Engineering and Technology, University of Technology and Applied Sciences-Salalah, Salalah 211, Sultanate of Oman*

*<sup>2</sup> Chemical Engineering, College of Engineering and Technology, University of Technology and Applied Sciences, Muscat 133, Sultanate of Oman*

*<sup>3</sup> Department of Chemical and Petroleum Engineering, School of Engineering & Computing, American University of Ras*

*Al Khaimah (AURAK), Ras al Khaimah 72603, United Arab Emirates*

*<sup>4</sup> Department of Production, OQ-Bi, Salalah 211, Sultanate of Oman*

### **ABSTRACT**

This study investigates the potential of Prosopis cineraria Leaves Powder (PCLP) as a biosorbent for removing lead (Pb) and zinc (Zn) from aqueous solutions, optimizing the process using Response Surface Methodology (RSM). Prosopis cineraria, commonly known as Khejri, is a drought-resistant tree with significant promise in environmental applications. The research employed a Central Composite Design (CCD) to examine the independent and combined effects of key process variables, including initial metal ion concentration, contact time, pH, and PCLP dosage. RSM was used to develop mathematical models that explain the relationship between these factors and the efficiency of metal removal, allowing the determination of optimal operating conditions. The experimental results indicated that the Langmuir isotherm model was the most appropriate for describing the biosorption of both metals, suggesting favorable adsorption characteristics.

#### \*CORRESPONDING AUTHOR:

Rakesh Namdeti, Chemical Engineering, College of Engineering and Technology, University of Technology and Applied Sciences-Salalah, Salalah 211, Sultanate of Oman; Email: rakesh.namdeti@utas.edu.om; Tel.: +968-9657-9136

#### ARTICLE INFO

Received: 10 October 2024 | Revised: 25 October 2024 | Accepted: 30 October 2024 | Published Online: 17 December 2024 DOI: https://doi.org/10.30564/jees.v7i1.7474

#### **CITATION**

Namdeti, R., Rao, G.B., Lakkimsetty, N.R., et al., 2024. Enhanced Lead and Zinc Removal via Prosopis Cineraria Leaves Powder: A Study on Isotherms and RSM Optimization. Journal of Environmental & Earth Sciences. 7(1): 292–305. DOI: https://doi.org/10.30564/jees.v7i1.7474

#### COPYRIGHT

Copyright © 2024 by the author(s). Published by Bilingual Publishing Group. This is an open access article under the Creative Commons Attribu tion-NonCommercial 4.0 International (CC BY-NC 4.0) License (https://creativecommons.org/licenses/by-nc/4.0/).

Additionally, the D-R isotherm confirmed that chemisorption was the primary mechanism involved in the biosorption process. For lead removal, the optimal conditions were found to be 312.23 K temperature, pH 4.72, 58.5 mg L<sup>-1</sup> initial concentration, and 0.27 g biosorbent dosage, achieving an 83.77% removal efficiency. For zinc, the optimal conditions were 312.4 K, pH 5.86, 53.07 mg L<sup>-1</sup> initial concentration, and the same biosorbent dosage, resulting in a 75.86% removal efficiency. These findings highlight PCLP's potential as an effective, eco-friendly biosorbent for sustainable heavy metal removal in water treatment.

*Keywords:* Prosopis Cineraria; Lead; Zinc; Isotherms; Optimization

### **1. Introduction**

Research has demonstrated that hazardous material releases and their environmental spread can have catastrophic consequences for people who are exposed<sup>[\[1\]](#page-12-0)</sup>. The enormous rise in the use of heavy metals over the last few decades has led to an increase in the flow of metallic materials into the aquatic environment<sup>[\[2\]](#page-12-1)</sup>. These metals' primary attribute is their persistence due to their non-degradable nature. Moreover, the majority of metal ions are harmful to living or-ganisms<sup>[\[3\]](#page-12-2)</sup>. Therefore, before wastewater is disposed of, harmful elements should be eliminated in order to provide a pollution-free environment.

Toxic metal ions can be extracted from aqueous solutions using a variety of techniques, such as chemical precipitation, ion exchange, membrane processes, reverse osmosis, solvent extraction, evaporation, and adsorption<sup>[\[4,](#page-12-3) [5\]](#page-12-4)</sup>. However, using the majority of these techniques has several drawbacks, such as the need for expensive machinery and space, the creation of sludge and other hazardous pollutants, and the need for electricity. These conventional methods could potentially have a number of drawbacks. For example, preparatory treatment may be necessary to eliminate complex compounds that could hinder precipitation. The minimal solubility of different metals in a mixed waste stream typically occurs at different pH values during precipitation. Therefore, the pH value at which the solubility of one element is too high may be the site of greatest removal of that metal. Furthermore, a successful precipitation process might not be the best way to address wastewater treatment issues given the increasingly strict treatment and disposal laws<sup>[\[6\]](#page-12-5)</sup>.

Many investigations have been conducted recently on the adsorption of heavy metals by various compounds. The development of non-traditional, low-cost adsorbents for the removal of metals from aqueous solutions, such as industrial and natural wastewater has been extensively documented.

Numerous studies have thoroughly examined the ability of different leaf materials. In the present study, the Leaves Powder of Prosopis cineraria were used to study the chemical form of the metal, solution pH, duration of contact, metal concentration, presence of competing adsorbates, amount of sorbent, organic matter, temperature, particle size, optimization methods, and otherf[act](#page-12-6)[or](#page-12-7)s that influence the adsorbability of dissolved elements[7–9] .

# **2. Methodology**

### **2.1. Sorbent**

The Prosopis cineraria leaves as showed in **Figure 1** from Oman's Dhofar Region, which were gathered from the Salalah region, were carefully processed. They were first thoroughly cleaned with distilled water to get rid of any impurities or debris. After being cleaned, the leaves were dried and ground into a fine powder using a home grinder. To guarantee complete moisture removal, the leaves were dried at 60 °C for 24 hours. Prior to usage, the dried leaves were sieved through a standard screen to produce a homogeneous particle size. The particle diameters varied between 75 and 212 μm.



**Figure 1.** Prosopis cineraria Leaves before and after drying.

#### **2.2. Chemicals**

Lead and zinc concentrations were generated by melting the relevant salts in distilled water and adjusting the pH with less hydrochloric acid and base.

#### **2.3. Experimentation**

To prepare the lead, and zinc solution, the pH value was reduced to 5.1 with HCl and NaOH, which were obtained by diluting the stock solution. To test the effect of agitation time on early lead and zinc levels, a mixture of 30 mL of metal solution and 0.1 g of 75 μm biosorbent was stirred at 180 rpm for a period of one hour. Centrifugation was performed at predetermined intervals, and the balance of lead and zinc in the solution was measured using AAS fitted with a hollow cathode and acetylene flame. Subsequent experiments used the same spectrophotometric approach. The amount of metal absorbed by Prosopis cineraria leaves was calculated using the following equation (Equation  $(1)$ ): the difference between the initial metal amount given to the powder material and the ion concentration for the supernatant.

$$
q = (C_0 - C_f) * (V/M) \tag{1}
$$

In the framework of metal adsorption, the parameter q  $(\text{mg g}^{-1})$  represents the weight of metal consumed per unit mass of adsorbent. The parameters  $C_0$  and  $C_f$  denote the metal's starting and ultimate concentrations, calculated in parts per million (ppm). V is the quantity of the solution in milliliters (mL), and M is the mass of the material that absorbs it in grams (g).

#### **2.4. Isotherms of Biosorption**

Biosorption isotherms explain how the absorbed metal ions and adsorbent material achieve equilibrium. Equilibrium isotherms are used to analyze the ability of Prosopis cineraria leaves to absorb lead and zinc metal ions. The Langmuir, Freundlich, Tempkin, and Dubinin-Radushkevich models are some of the most commonly employed to characterize such systems. These models are useful tools for analysing the kinetics of metal ion adsorption on biosorbents such as Prosopis cineraria leaves.

#### **2.5. CCD**

Using full factorial rotatable central composite design (CCD), the impact of several process factors, including temperature  $(X1)$ , solution pH  $(X2)$ , biosorbent dosage  $(X3)$ , and initial metal concentration (X4), on the removal of lead and zinc metal from aqueous solutions was investigated  $[10]$ . Thirty experiments were needed in total: sixteen cube point runs, eight axial point runs, and two center point runs. Each experiment had to be completed in duplicate. Every experiment was run with a constant agitation speed of 180 rpm and a contact time (t).

Using Equation  $(2)^{[11]}$  $(2)^{[11]}$  $(2)^{[11]}$ , all independent variables were coded to five levels as Xi.

$$
\mathbf{x}_{\mathbf{i}} = \frac{X_i - X_{oi}}{\Delta X_i}, \ \mathbf{i} = 0, 1, 2, 3, \dots \mathbf{k} \tag{2}
$$

Where  $x_i$  is the real value of an independent variable, x0i is the real value of the independent variable at the center point,  $\Delta x_i$  is the step change, and  $X_i$  is the dimensionless value of an independent variable. Using STATISTICA 6.0 (Stat Soft Inc.), a second-degree polynomial equation (Equation (3)) was created to calculate the percentage of metal biosorption at various biosorption process operating conditions.

$$
Y = b_0 + b_1 X_1 + b_2 X_2 + b_3 X_3 + b_4 X_4
$$
  
+ $b_{11}X_1^2 + b_{22}X_2^2 + b_{33}X_3^2 + b_{44}X_4^2$   
+ $b_{12}X_1X_2 + b_{13}X_1X_3 + b_{14}X_1X_4$   
+ $b_{23}X_2X_3 + b_{24}X_2X_4 + b_{34}X_3X_4$  (3)

Where, Y is the predicted response,  $X_1$ ,  $X_2$ ,  $X_3$  and  $X_4$ are independent variables:  $b_0$  is an offset term;  $b_1$ ,  $b_2$ ,  $b_3$ and  $b_4$  are linear effects;  $b_{11}$ ,  $b_{22}$ ,  $b_{33}$  and  $b_{44}$  are squared effects; and  $b_{12}$ ,  $b_{13}$ ,  $b_{14}$ ,  $b_{23}$ ,  $b_{24}$  and  $b_{34}$  are interaction terms.

### **3. Results**

#### **3.1. Isotherms for Biosorption**

#### **3.1.1. Isotherm of Langmuir**

This sorption model, given by him in reference  $[12]$ , served to estimate the biosorbent's maximal ion sorpt[ion](#page-12-10) capacity. This isotherm may be stated as follows in Equation  $(4)$ :

$$
q_{eq} = \frac{q_{\text{max}} K_a C_{eq}}{1 + K_a C_{eq}} \tag{4}
$$

To fit the experimental results, the Langmuir model was linearized, with " $q_{\text{max}}$ " representing the biosorbent's monolayer biosorption capacity measured in mg  $g^{-1}$  and the Langmuir constant " $K_a$ " (L mg<sup>-1</sup>) corresponding to the energy of biosorption as given in the Equation (5).

$$
\frac{C_{eq}}{q_{eq}} = \frac{C_{eq}}{q_{\text{max}}} + \frac{1}{K_a q_{\text{max}}}
$$
(5)

The relationship of sorption  $(C_{eq}/q_e)$  versus concentration  $(C_{ea})$  show this removal follows Langmuir, as shown in **Figures 2** and **3** (Equation (5)). Experimental data for lead, zinc ion sorption on PCLP at various temperatures confirm this conclusion. **Tables 1** and **2** shows the Langmuir constants: saturated monolayer sorption capacity  $(q_{max})$  and sorption stability constant  $(K_a)$  for lead and zinc ion sorption onto PCLP at 303, 308, 313, and 318 K. The  $(R^2)$  values clearly imply the lead, zinc ion sorption on PCLP follows isotherm. Notably, the maximal ion removal  $(q<sub>max</sub>)$  was found to be 31.98 mg  $g^{-1}$  and 27.66 mg  $g^{-1}$  respectively, which gradually decreased as the temperature increased.



**Figure 2.** The Langmuir isotherm of lead adsorption by the PCLP biosorbent at different temperatures.



**Figure 3.** Langmuir isotherm for zinc by PCLP biosorbent at different temperatures.

#### **3.1.2. Freundlich**

The isotherm of Freundlich<sup>[\[13\]](#page-12-11)</sup> as given in Equation (6):

$$
q_e = K_F C_{eq}^{\frac{1}{n}} \tag{6}
$$

The formula can be transformed by using logarithmic values to determine the variables  $K_F$  and n as shown in Equation (7):

$$
\ln q_e = \ln K_F + \frac{1}{n} \ln C_{eq} \tag{7}
$$

The constants were shown in **Tables 1** and **2**. **Figures 4** and **5** presents an exponential equation (Equation (7)).



**Figure 4.** The Freundlich isotherm of lead adsorption by the PCLP biosorbent at different temperatures.



**Figure 5.** Freundlich isotherm for zinc by PCLP biosorbent at different temperatures.

### **3.1.3. Temkin**

Tempkin<sup>[\[14\]](#page-12-12)</sup> examined thermal properties for biosorption and the effect of adsorbate on biosorption isotherms. They argued that due to those interactions, the biosorption energy for all compounds falls linearly with coverage (**Figures 6** and **7**).

The Tempkin isotherm has been applied in the form below as given in Equation (8).

$$
q_{eq} = \frac{RT}{b} \ln(A_T C_{eq})
$$
\n(8)

And linearized as shown in Equation (9):

$$
q_{eq} = B \ln A_T + B \ln C_{eq} \tag{9}
$$



**Figure 6.** The Tempkin isotherm of lead adsorption by the PCLP biosorbent at different temperatures.



**Figure 7.** Tempkin isotherm for zinc by PCLP biosorbent at different temperatures.

### **3.1.4. Dubinin-Radushkevich**

The equilibrium interaction between biosorption and adsorbate with a specific metal-powder partnership could be stated irrespective of temperature by using the removal ability  $(\varepsilon)$ , as defined in Equation (10).

$$
\varepsilon = RT(1 + 1/C_{\text{eq}}) \tag{10}
$$

The isotherm adopts a distinctive curve, and it is represented using Equation (11).

$$
\ln q_e = \ln Q_D - K\varepsilon^2 \tag{11}
$$

This may be derived using (**Figures 8** and **9**) Equation (12):

$$
E = 1/\sqrt{2K} \tag{12}
$$

Drawing ln qe vs  $\varepsilon^2$ , the Q<sub>D</sub> & K are determined.



**Figure 8.** The Dubinin-Radushkevich isotherm of lead adsorption by the PCLP biosorbent at different temperatures.



**Figure 9.** Dubinin-Radushkevich isotherm for zinc by PCLP biosorbent at different temperatures.

Analysis of **Tables 1** and **2** reveals that the adsorption of lead, zinc on the PCLP exhibited strong correlations with different isotherms.

The equilibrium data were tested by all the four isotherm models. Based on the high correlation coefficient  $(R<sup>2</sup>)$  values presented in **Tables 1** and 2 the best isotherm was determined. From these values, it was observed that Langmuir Isotherm with correlation coefficient of 0.998 fitted well followed by Freundlich, Tempkin and D-R isotherms with a correlation coefficient of 0.988, 0.969 and 0.843, respectively at a solution temperature of 303 K. For Zinc removal, the biosorption data was well represented by Langmuir model with correlation coefficient of 0.9943 followed by Freundlich, Tempkin and D-R isotherms with the correlation coefficients

Temperature T(K)	Langmuir			Freundlich			Tempkin			<b>Dubinin Radushkevich</b>			
	$Ka$ , $(L)$ $mg^{-}$	qmax (mg	$R^2$	Kf, (mg $-$	п	ີ $R^2$	AT (L $mg^{-}$			QD (mg) $-$	$mol2KJ-2$	E (KJ $mol-1$	$R^2$
304 308 313 318	0.06 0.18 0.33 0.65	31.9 26.93 25.90 25.57	0.99 0.996 0.990 0.991	2.86 5.67 7.38 9.72	l.62 2.28 2.56 2.94	0.988 0.992 0.962 0.988	0.667 2.344 4.392 10.382	354.3 485.9 524.4 575.6	0.969 0.965 0.992 0.969	17.881 18.037 19.735 20.731	0.0048 0.0091 0.0060 0.0034	10.1488 7.3779 9.0825 12.0554	0.843 0.802 0.896 0.959

**Table 1.** Bisorption isotherm constants for lead removal with PCLP biosorbent.

**Table 2.** Bisorption isotherm constants for zinc removal using PCLP biosorbent.

		<b>Langmuir Isotherm</b>		<b>Freundlich Isotherm</b>			<b>Tempkin Isotherm</b>			Dubinin Radushkevich Isotherm			
Temperature T(K)	qmax (mg g	Ka, (L mg	$\mathbb{R}^2$	n	Kf, (mg -	$R^2$	b	AT(L) mg	$R^2$		qmax (mg g	$R^2$	E (KJ $mol-1$
303 308 313 318	27.66 22.41297 21.12995 21.38628	0.060 0.133376 0.266304 0.573465	0.9943 0.99808610 0.99554364 0.99831975	1.721 2.2679 2.882413 3.500519	2.535 4.14251 6.084599 8.424316	0.978 0.9740223 0.9331284 0.8678837	412.18 536.73 658.78 758.81	0.59370 1.39623 3.94492 13.1665	0.99456 0.99970 0.98770 0.95686	0.00802624 0.00289489 0.01150483 0.00459704	15.82393 15.5751 16.51124 18.44433	0.90332 0.88372 0.92702 0.9728	8.89276 13.1422 6.59242 10.4291

of 0.9785, 0.99 and 0.903 respectively at a solution temperature of 303 K. At remaining temperatures also, the same type of trends was observed for both the metals.

# **4. Optimization of the Selected Biosorption Process Parameters Using RSM**

It is extremely challenging to describe and simulate the biosorption process due to its high level of complexity and the interplay of numerous process factors. Although the one factor at a time approach is straightforward, it is unable to capture the interplay between all the variables that contribute to the biosorption process. It also necessitates a lengthy process and numerous experimental runs. Consequently, several processes were developed, enhanced, and optimized using the Response Surface Methodology (RSM). It provides details on the full interactions between all process parameters that have an impact on the process  $[15-18]$  $[15-18]$ . The percentage of metal biosorption and the metal absorption capacity of PCLP leaves biosorbent were found to be highly influenced by solution temperature, solution pH, initial concentration of metal, and biosorbent dosage, according to an analysis of preliminary experimental results. Thus, using full factorial rotatable central composite design (CCD), these four process variables were chosen to ascertain the ideal process parameters for the greatest removal of Lead and Zinc onto PCLP leaves biosorbent. **Table 3** lists the levels and ranges of the selected independent factors that were employed in the metal removal studies.

To optimize the parameters, a 24-factorial CCD design

was utilized, consisting of eight axial points ( $\alpha = -4$ ) and six replications at the center points ( $no = 6$ ). This resulted in a total of thirty experiments (**Table 4** for lead and **Table 5** for zinc). The % elimination of lead and zinc  $(Y)$  was found to be a function of temperature  $(X1)$ , pH  $(X2)$ , biosorbent dosage  $(X3)$ , and initial concentration  $(X4)$ , according to the computed regression Equations (13) and (14) for the optimization of medium ingredients. Multiple regression analysis was performed on the experimental data, resulting in the following equations for lead (Y1) and zinc (Y2) biosorption:

$$
Y_1(\text{\% Biosorption of Lead}) = -333.349 + 21.237X_1 - 7.397X_2 + 17.138X_3 + 0.535X_4 - 0.299X_1^2 - 2.209X_2^2 - 145.938X_3^2 - 0.007X_4^2 + 0.516X_1X_2 - 0.035X_1X_3 - 0.003X_1X_4 + 12.8X_2X_3 + 0.078X_2X_4 - 0.051X_3X_4
$$
  
\n
$$
Y_2(\text{\% Biosorption of Zinc}) = -212.740
$$

+12.065X<sup>1</sup> + 1.747X<sup>2</sup> + 67.310X<sup>3</sup> + 0.481X<sup>4</sup> –0.171X<sup>1</sup> 2 –1.954X<sup>2</sup> <sup>2</sup> + 34.332X<sup>3</sup> 2 –0.003X<sup>4</sup> 2 +0.574X1X2–2.085X1X3–0.005X1X<sup>4</sup> +0.949X2X<sup>3</sup> + 0.007X2X4–0.118X3X<sup>4</sup> (14)

For lead and zinc, the regression model's coefficients were determined and are shown in **Tables 6** and **7**. They contain one block term, four linear, four quadratic and six interaction terms. The significance of each coefficient was determined by student's *t*-test and *p*-values and listed in **Table 6** for lead and **Table 7** for zinc. The relevant coefficient was more significant the bigger the t-value's magnitude and the smaller the p-value. As can be seen from their individual p-values (**Table 6** for lead and **Table 7** for zinc), this suggests that the linear, quadratic, and interaction effects of temperature, pH, biosorbent dosage, and initial concentration of

lead are very significant. The parity plot (**Figure 10** for lead and **Figure 11** for zinc) showed a satisfactory correlation between the experimental and predicted values of percentage removal of lead indicating good agreement of model data with the experimental data.



**Figure 10.** A parity plot that displays the distribution of the actual versus projected lead biosorption % using PCLP.



**Figure 11.** A parity plot that displays the distribution of the actual against anticipated values for the zinc biosorption percentage using PCLP.

**Table 8** displays the findings of the ANOVA-fitting second-order response surface model for lead and zinc. An accurate statistical tool for evaluating the model's significance and suitability is the Fisher variance ratio, or F-value  $( = SR<sup>2</sup>/Se<sup>2</sup>)$ . The likelihood that the factors can sufficiently explain the variance in the data about its mean and that the estimated factor effects are real increases with the F-value above unity. Regression model ANOVA revealed an extremely low probability value (Pmodel  $> F = 0.000000$ ) and a very significant model, as evidenced by the Fisher's F-test  $(Fmodel = 374.11$  for Lead 718.5 for Zinc).

Furthermore, the F-value (F0.05(14.15) =  $SR^2/Se^2$  = 374.11 for Lead and 718.5 for Zinc) computed at the 1% level was greater than the F-value (F0.05 (14.15) tabulars  $=$ 2.46 for both metals, demonstrating the significance of the treatment differences. An indicator of the model's variability in the observed response values is the correlation coefficient  $(R<sup>2</sup>)$ . The stronger and more accurate the model's response prediction, the closer the  $R^2$  value is to 1. The correlation coefficient values in this investigation ( $R^2 = 0.9979$  for lead and 0.9985 for zinc) showed that the model could account for 99.79 percent (for lead) and 99.85% (for zinc) of the variability in the response. To further support the model's high significance, the adjusted correlation coefficient values (Adj  $R^2 = 0.9952$  for lead and 0.9971 for zinc) were also extremely high<sup>[\[19](#page-13-1)[–23\]](#page-13-2)</sup>.

**Figures 12–17** for lead and **Figures 18–23** for zinc display the response surface plots of the percentage of lead biosorption vs the interactive effect of pH, beginning lead concentration, biosorbent dosage, and temperature. The 3D response surface graphs in **Figures 12−17** for Lead and **Figures 18−23** for Zinc visually represented regression Equations (13) and (14). By holding the values of the other variables constant, the response surface graphs display the relative effects of any two variables. **Figure 12**'s response surface plot illustrates how temperature and solution pH interact when starting metal concentration and biosorbent dosage are maintained at midpoints.In a similar vein, the subsequent figures describe how various parameters interact to affect the proportion of metal biosorption<sup>[\[24](#page-13-3)[–26\]](#page-13-4)</sup>. With one test parameter kept at zero, each response plot shows various combinations of the other two test parameters. The surface confined in the smallest curve (elliptical or circular) of the response plot represents the maximum percentage of biosorption of the level. Aqueous solution initial metal concentrations of 58.523 mg L<sup>-1</sup> for lead and 53.07 mg L<sup>-1</sup> for zinc, biosorbent dosages of 0.271 g  $L^{-1}$  for both lead and zinc, pH values of 4.72 for lead and 5.86 for zinc, and temperatures of 39.230C for lead and 39.420C for zinc are the ideal set of parameters for maximizing the percentage of biosorption of lead and zinc. Under these ideal circumstances, the percentage of lead and zinc that underwent biosorption was 83.77 and 75.66%, respectively. **Table 8**[\[27–](#page-13-5)[29\]](#page-13-6) displays the ideal variable values for lead and zinc biosorption derived from the regression equation.



**Table 3.** Range and concentrations of the independent parameters under experimentation for the biosorption of lead and zinc onto PCLP.

Experiments carried out under these ideal circumstances validated these conditions. The greatest percentage of biosorption for lead and zinc at these ideal conditions was found to be 80.23% and 72.31%, respectively. These experimental data clearly show that the experimental results correlate well with the outcomes predicted by the RSM. These findings also shown the effectiveness of applying statistical experimental designs utilizing RSM to optimize different parameters in order to enhance lead and zinc biosorption onto PCLP leaves biosorbent. These findings showed that a particular combination of temperature, solution pH, biosorbent dosage, and starting metal concentration led to the greatest percentage of metal biosorption.

Although this study shows that Prosopis cineraria Leaves Powder (PCLP) is a good biosorbent for removing zinc and lead, it should be noted that it has some limits. First off, the intricacies of continuous-flow systems, which are more prevalent in real-world applications, are not adequately captured by the batch trials carried out here. The scalability of PCLP biosorption in dynamic systems may be examined in future research. Additionally, real-world wastewater may contain a variety of pollutants that could impact biosorption effectiveness, even if the response surface methodology (RSM) made it possible to optimize process parameters.A more thorough grasp of PCLP's uses might be possible by investigating how well it performs in situations with several metals and organic contaminants. Lastly, even though the D-R isotherm indicated chemisorption as the main mechanism, additional molecular research might validate these interactions and investigate changes to improve the capacity and reusability of PCLP. According to the study's findings, PCLP has potential as an affordable, environmentally responsible method of treating water sustainably and supporting international initiatives to lessen heavy metal contamination. Pilot-scale testing and industry partnerships will be the next stages in incorporating this biosorbent into workable wastewater treatment systems.



**Figure 12.** Response surface plot showing how temperature and pH affect the amount of lead that PCLP is able to biosorb.



**Figure 13.** RSM plot showing how temperature and biosorbent dosage affect the proportion of lead that PCLP biosorbs.



**Figure 14.** RSM plot showing how temperature and starting metal concentration affect the percentage of lead that PCLP biosorbs.

		<b>Values with Codes</b>				<b>Actual Numbers</b>		% Sorption of Lead		
Run No.	x1	x2	x3	x4	X1	X2	X3	<b>X4</b>	<b>Noted</b>	<b>Expected</b>
$\mathbf{1}$	$-1$	$-1$	$-1$	$-1$	35	$\overline{4}$	0.2	40	77.13	77.26292
$\overline{c}$	$-1$	$-1$	$-1$	$\mathbf{1}$	35	4	0.2	80	74.85	74.53125
3	$-1$	$-1$	$\mathbf{1}$	$-1$	35	4	0.4	40	73.06	73.58292
4	$-1$	$-1$	$\mathbf{1}$	$\mathbf{1}$	35	4	0.4	80	71.13	71.26125
5	$-1$	1	$-1$	$-1$	35	6	0.2	40	65.67	65.75125
6	$-1$	1	$-1$	$\mathbf{1}$	35	6	0.2	80	68.99	69.24958
7	$-1$	1	1	$-1$	35	6	0.4	40	67.59	67.19125
$\,$ $\,$	$-1$	1	$\mathbf{1}$	$\mathbf{1}$	35	6	0.4	80	71.19	71.09959
9	1	$-1$	$-1$	$-1$	45	4	0.2	40	69.15	69.50792
10	1	$-1$	$-1$	$\mathbf{1}$	45	4	0.2	80	64.97	65.62625
11		$-1$	$\mathbf{1}$	$-1$	45	4	0.4	40	65.76	65.75792
12		$-1$	$\mathbf{1}$	$\mathbf{1}$	45	4	0.4	80	62.1	62.28625
13		$\mathbf{1}$	$-1$	$-1$	45	6	0.2	40	68.19	68.31625
14		1	$-1$	$\mathbf{1}$	45	6	0.2	80	70.92	70.66458
15	1	1	$\mathbf{1}$	$-1$	45	6	0.4	40	69.1	69.68625
16	1	1	1	$\mathbf{1}$	45	6	0.4	80	72.32	72.44458
17	$-2$	$\mathbf{0}$	$\mathbf{0}$	$\mathbf{0}$	30	5	0.3	60	56.7	56.8025
18	$\overline{c}$	$\mathbf{0}$	$\mathbf{0}$	$\mathbf{0}$	50	5	0.3	60	51.02	50.3925
19	$\mathbf{0}$	$-2$	$\overline{0}$	$\mathbf{0}$	40	3	0.3	60	75.95	75.37917
20	$\mathbf{0}$	$\overline{c}$	$\mathbf{0}$	$\mathbf{0}$	40	$\overline{7}$	0.3	60	73.98	74.02583
21	$\theta$	$\mathbf{0}$	$-2$	$\mathbf{0}$	40	5	0.1	60	78.91	78.6525
22	$\theta$	$\mathbf{0}$	$\overline{c}$	$\theta$	40	5	0.5	60	77.02	76.7525
23	$\mathbf{0}$	$\mathbf{0}$	$\boldsymbol{0}$	$-2$	40	5	0.3	20	72.98	72.53917
24	$\theta$	$\theta$	$\mathbf{0}$	$\overline{c}$	40	5	0.3	100	72.65	72.56583
25	$\theta$	$\mathbf{0}$	$\mathbf{0}$	$\mathbf{0}$	40	5	0.3	60	83.54	83.54
26	$\mathbf{0}$	$\mathbf{0}$	$\mathbf{0}$	$\mathbf{0}$	40	5	0.3	60	83.54	83.54
27	$\mathbf{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	$\mathbf{0}$	40	5	0.3	60	83.54	83.54
28	$\theta$	$\mathbf{0}$	$\mathbf{0}$	$\mathbf{0}$	40	5	0.3	60	83.54	83.54
29	$\theta$	$\boldsymbol{0}$	$\mathbf{0}$	$\mathbf{0}$	40	5	0.3	60	83.54	83.54
30	$\theta$	$\overline{0}$	$\overline{0}$	$\theta$	40	5	0.3	60	83.54	83.54

**Table 4.** The central composite design matrix displays both real and coded values for the proportion of lead biosorption with PCLP, in addition to the observed and anticipated values.

Note:  $X_1$  = Temperature,  $X_2$  = pH,  $X_3$  = Biosorbent dosage,  $X_4$  = Initial metal concentration.



**Figure 15.** RSM graphic showing how pH and biosorbent dosage affect the percentage of lead that PCLP is able to biosorb.



**Figure 16.** RSM plot showing how pH and starting metal concentration affect the percentage of lead that PCLP is able to biosorb.

			<b>Values with Codes</b>			<b>Actual Numbers</b>		% Sorption of Zinc		
Run No.	x1	x2	x3	x4	X1	X2	X3	<b>X4</b>	<b>Noted</b>	<b>Expected</b>
$\mathbf{1}$	$-1$	$-1$	$-1$	$-1$	35	5	0.2	40	69.2	69.50166
$\overline{c}$	$-1$	$-1$	$-1$	$\mathbf{1}$	35	5	0.2	80	58.38	58.39208
3	$-1$	$-1$	$\mathbf{1}$	$-1$	35	5	0.4	40	66.13	66.44041
4	$-1$	$-1$	$\mathbf{1}$	$\mathbf{1}$	35	5	0.4	80	59.66	59.52833
5	$-1$	1	$-1$	$-1$	35	7	0.2	40	62.22	62.48542
6	$-1$	$\mathbf{1}$	$-1$	$\mathbf{1}$	35	7	0.2	80	60.12	60.19333
$\sqrt{ }$	$-1$	$\mathbf{1}$	$\mathbf{1}$	$-1$	35	$\sqrt{ }$	0.4	40	59.01	59.06167
$\,$ 8 $\,$	$-1$	$\mathbf{1}$	$\mathbf{1}$	1	35	7	0.4	80	60.23	60.96708
9	1	$-1$	$-1$	$-1$	45	5	0.2	40	65.92	65.58542
10	1	$-1$	$-1$	$\mathbf{1}$	45	5	0.2	80	60.21	60.40833
11	1	$-1$	$\mathbf{1}$	$-1$	45	5	0.4	40	63.2	63.37667
12		$-1$	$\mathbf{1}$	$\mathbf{1}$	45	5	0.4	80	62.26	62.39708
13		1	$-1$	$-1$	45	7	0.2	40	59.29	59.67167
14		1	$-1$	$\mathbf{1}$	45	7	0.2	80	63.22	63.31208
15	1	1	$\mathbf{1}$	$-1$	45	$\sqrt{ }$	0.4	40	56.71	57.10042
16	1	1	$\mathbf{1}$	$\mathbf{1}$	45	$\sqrt{ }$	0.4	80	64.99	64.93833
17	$-2$	$\mathbf{0}$	$\boldsymbol{0}$	$\mathbf{0}$	30	6	0.3	60	65.98	65.49625
18	$\overline{c}$	$\mathbf{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	50	6	0.3	60	65.72	65.55125
19	$\mathbf{0}$	$-2$	$\boldsymbol{0}$	$\boldsymbol{0}$	40	4	0.3	60	49.12	49.11125
20	$\mathbf{0}$	$\boldsymbol{2}$	$\mathbf{0}$	$\mathbf{0}$	40	8	0.3	60	45.28	44.63625
21	$\boldsymbol{0}$	$\mathbf{0}$	$-2$	$\mathbf{0}$	40	6	0.1	60	69.98	69.81125
22	$\boldsymbol{0}$	$\mathbf{0}$	$\overline{c}$	$\mathbf{0}$	40	6	0.5	60	68.86	68.37625
23	$\boldsymbol{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	$-2$	40	6	0.3	20	68.93	68.48458
24	$\boldsymbol{0}$	$\mathbf{0}$	$\boldsymbol{0}$	$\overline{c}$	40	6	0.3	100	65.42	65.21291
25	$\boldsymbol{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	40	6	0.3	60	75.82	75.82
26	$\mathbf{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	40	6	0.3	60	75.82	75.82
27	$\mathbf{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	$\boldsymbol{0}$	40	6	0.3	60	75.82	75.82
28	$\theta$	$\boldsymbol{0}$	$\boldsymbol{0}$	$\theta$	40	6	0.3	60	75.82	75.82
29	$\boldsymbol{0}$	$\boldsymbol{0}$	$\mathbf{0}$	$\boldsymbol{0}$	40	6	0.3	60	75.82	75.82
30	$\theta$	$\mathbf{0}$	$\overline{0}$	$\mathbf{0}$	40	6	0.3	60	75.82	75.82

**Table 5.** The central composite design matrix displays both real and coded values for the proportion of zinc biosorption with PCLP, in addition to the observed and anticipated values.

Note:  $X_1$  = Temperature,  $X_2$  = pH,  $X_3$  = Biosorbent dosage,  $X_4$  = Initial metal concentration.



**Figure 17.** RSM map showing how the dosage of the biosorbent and the starting metal concentration affect the percentage of lead that is biosolated by PCLP.



**Figure 18.** Response surface map showing how temperature and starting metal concentration affect PCLP's zinc biosorption %.



Note:  $X_1$  = Temperature,  $X_2$  = pH,  $X_3$  = Biosorbent dosage,  $X_4$  = Initial metal concentration.<br><sup>a</sup> Significant (p ≤ 0.05), R<sup>2</sup> = 0.9979; Adj R<sup>2</sup> = 0.9952.





Note:  $X_1$  = Temperature,  $X_2$  = pH,  $X_3$  = Biosorbent dosage,  $X_4$  = Initial metal concentration.<br><sup>a</sup> Significant (p ≤ 0.05), R<sup>2</sup> = 0.9985; Adj R<sup>2</sup> = 0.9971.





**Figure 19.** RSM plot showing how the pH and starting metal concentration affect the proportion of zinc that PCLP is able to biosorb.

**Figure 20.** Response surface plot of the effects of initial metal concentration and biosorbent dosage on percentage of biosorption of zinc by PCLP.







ture on percentage of biosorption of zinc by *Grewia Orbiculata* L.



**Figure 22.** Response surface plot of the effects of biosorbent dosage and temperature on percentage of biosorption of zinc by PCLP.



**Figure 23.** RSM graphic showing how pH and biosorbent dosage affect PCLP's zinc biosorption %.

## **5. Conclusions**

The results of the experiment showed that the Langmuir isotherm model accurately characterized the biosorption of lead and zinc onto Prosopis cineraria Leaves Powder (PCLP), suggesting advantageous adsorption properties. The Dubinin-Radushkevich (D-R) isotherm's E values further revealed that the biosorption process primarily took place in the chemisorption region, just over the physisorption threshold. Response Surface Methodology (RSM) with a central composite design (CCD) was applied to optimize temperature, solution pH, biosorbent dosage, and initial metal ion concentration, among other important experimental parameters. The model's dependability was validated by the near alignment of the optimal conditions predicted by the RSM with the experimental data. At a temperature of 312.23 K, pH of 4.72, initial metal concentration of 58.5 mg  $L^{-1}$ , and biosorbent dosage of 0.27 g, the best removal efficiency of 83.77% was obtained for lead. Similar predictions were made for zinc, indicating a maximum removal efficiency of 75.86% at 312.4 K, pH 5.86, 53.07 mg L−<sup>1</sup> beginning metal concentration, and the same biosorbent dosage. These findings highlight PCLP's potential as an efficient biosorbent for heavy metal removal, offering a viable, long-term option for water treatment and environmental remediation.

### **Author Contributions**

The authors of this study made the following contributions: R.N. conceptualized the research, designed the study, supervised the process, and contributed to data analysis and manuscript writing. G.B.R. conducted the experimental work; including the preparation of Prosopis cineraria Leaves Powder (PCLP) and performing biosorption experiments for lead and zinc removal. N.R.L. assisted in experimental design, data collection, analysis, and statistical modeling. N.M.S.Q. provided support in data analysis, result interpretation, and manuscript editing. N.P.B.S. performed statistical analysis using Response Surface Methodology (RSM) and contributed to mathematical model development. U.R.M. supported the experimental work, characterization of biosorbents, and manuscript writing. P.P.M. assisted with laboratory work, data collection, and interpretation of results. D.S.M.S.A.-K. contributed to the manuscript writing, especially discussing the ecological implications of using PCLP as a biosorbent. M.A.A.Q. provided insights into adsorption mechanisms and supported the manuscript's discussion. H.A.S.B.A. coordinated the study, reviewed results, and contributed to drafting and finalizing the manuscript.

# **Funding**

This work received no external funding.

## **Institutional Review Board Statement**

Not applicable.

### **Informed Consent Statement**

Not applicable.

### **Data Availability Statement**

The data that support the findings of this study are available from the corresponding author upon reasonable request. All relevant data are included in the manuscript and its supplementary materials.

# **Conflicts of Interest**

The authors declare no conflict of interest.

# **References**

- <span id="page-12-0"></span>[1] Benjamin, M.M., Hayes, K.F., Leckie, J.O., 1982. Removal of toxic Metals from power generation waste streams by Adsorption and co-precipitation. Journal of the Water Pollution Control Federation. 54(11), 1472–1481.
- <span id="page-12-1"></span>[2] Namasivayam, C., Ranganathan, K., 1995. Removal of Cd (II) from wastewater by adsorption on waste Fe(III)/Cr(III) hydroxide. Water Research. 29(7), 1737–1744.
- <span id="page-12-2"></span>[3] Budinova, T.K., Gergova, K.M., Petrov, N.V., et al., 1995. Removal of metal ions from aqueous solution by activated carbons obtained from different raw materials. Journal of Chemical Technology & Biotechnology. 60, 177–182.
- <span id="page-12-3"></span>[4] Srivastava, S.K., Gupta, V.K., Mohan, D., 2000. Removal of lead and chromium by activated slag—a blastfurnace waste. Journal of Environmental Engineering. 123(5), 461–468.
- <span id="page-12-4"></span>[5] Zouboulis, A.L., Matis, K.A., 1997. Removal of metal ions from dilute solutions by sorption flotation. Critical Review in Environmental. 27(3), 195–235.
- <span id="page-12-5"></span>[6] Chen, W.J., Anderson, P.R., Holsen, T.M., 1991. Recovery and recycle of metals from wastewater with magnetite-based adsorption process. Research Journal WPCF. 63(7), 954–964.
- <span id="page-12-6"></span>[7] Orumwense, F.F.O., 1996. Removal of lead from water by adsorption on a kaolinitic clay. Journal of Chemical Technology & Biotechnology. 65, 363–369.
- [8] Johnson, B., 1990. Effect of pH, temperature and concentration on the adsorption of cadmium on goethite. Environmental Science & Technology. 24, 112–118.
- <span id="page-12-7"></span>[9] Farrah, H., Pickering, W.F., 1977. The sorption of lead and cadmium species by clay minerals. Australian Journal of Chemistry. 30, 1414–1422.
- <span id="page-12-8"></span>[10] Babak, L., Supinova, P., Zichova, M., et al., 2012. Biosorption of Cu, Zn and Pb by thermophilic bacteria – effect of biomass concentration on biosorption capacity. Acta Universitatis Agriculturae et Silviculturae Mendelianae Brunensis. 5, 55–59.
- <span id="page-12-9"></span>[11] Elliott, H.A., Huang, C.P., 1981. Adsorption characteristics of some Cu(II), Pb(II) complexes on alluminosilicates. Water Research. 15, 849–855.
- <span id="page-12-10"></span>[12] Galun, M., Galun, E., Siegel, B.Z., et al., 1987. Removal of metal ions from aqueous solutions by Penicillium biomass: Kinetic and uptake parameters. Water, Air & Soil Pollution. 33, 359–371.
- <span id="page-12-11"></span>[13] Esteves, A.J.P., Valdman, E., Leite, S.G.F., 2000. Repeated removal of cadmium and zinc from an industrial effluent by waste biomass Sargassum sp. Biotechnology Letters. 22, 499–502.
- <span id="page-12-12"></span>[14] Hall, K.R., Eagleton, L.C., Acrivos, A., et al., 1966. Pore- and diffusion kinetics in fixed bed adsorption under constant-pattern conditions. Industrial & Engineering Chemistry Fundamentals. 5(2), 212–223.
- <span id="page-12-13"></span>[15] Kara, S., Aydiner, C., Demirbas, E., et al., 2007. Modelling the effects of adsorbent dose and particle size on the adsorption of reactive textile dye by fly ash. Desalination. 212(1–3), 282–293.
- [16] Shi, Y., Xu, X., Zhu, Y., 2009. Optimization of verticillium lecanii spore production in solid state fermentation on sugarcane bagasse. Applied Microbial Biotechnology. 82, 921–927.
- [17] Fazil, M.M., Medaghinia, A., Naddafi, K., et al., 2010. Optimization of rice reactivate blue 19 decolorization

by Ganoderma sp. Using response surface methodology. Iranian Journal of Environmental Health Science & Engineering. 7(1), 35–42.

- <span id="page-13-0"></span>[18] Myers, R.H., Montgomery, D.C., 2002. Response surface methodology: Process and product optimization using designed experiments. Wiley: New York, USA. pp. 856-894.
- <span id="page-13-1"></span>[19] King, P., Rakesh, N., Beenalahari, S., et al., 2007. Removal of lead from aqueous solution using *Syzygium cumini* L.: Equilibrium and kinetic studies. Journal of Hazardous Materials. 142(1–2), 340–347.
- [20] King, P., Rakesh, N., Lahari, S.B., et al., 2008. Biosorption of zinc onto *Syzygium cumini* L.: Equilibrium and kinetic studies. Chemical Engineering Journal. 144(2), 181–187.
- [21] Tadepalli, S., Murthy, K.S.R., Rakesh, N.N., 2016. Removal of Cu (II) and Fe (II) from Industrial waste water using orange peel as adsorbent in batch mode operation. International Journal of ChemTech Research. 9(5), 290–299.
- [22] Rakesh, N., Kalpana, P., Rao, L.N., et al., 2010. Removal of zinc ions from aqueous solution by *Ficus Benghalensis* L.: Equilibrium and kinetic studies. International Journal of Engineering Studies. 2(1), 15–28.
- <span id="page-13-2"></span>[23] Al-Amri, L.S.A.A., Al-Subhi, M.H.H., Namdeti, R., 2014. Comparison studies for the removal of methylene blue from aqueous solution using Tea and Coffee

powder. International Journal of ChemTech Research. 6(1), 619–627.

- <span id="page-13-3"></span>[24] Namdeti, R., Pulipati, K., 2014. Lead removal from aqueous solution using Ficus Hispida leaves powder. Desalination and Water Treatment. 52(1–3), 339–349.
- [25] Namdeti, R., 2023. A review on removal of heavy metals by biosorption: A green technology. International Journal of Research and Review. 10(3), 531–543.
- <span id="page-13-4"></span>[26] Kalpana, P., Rakesh, N., Nageswara, R.L., 2010. Biosorption of lead ions from aqueous solution by Terminalia catappa: Equilibrium and kinetic studies. Nature Environment and Pollution Technology. 9(2), 335–343.
- <span id="page-13-5"></span>[27] Joaquin, A., Al Hadrami, S.H.A., Namdeti, R., 2015. Water analysis using activated carbon from coconut shell. International Journal of Latest Research in Science and Technology. 4(5), 1–3.
- [28] Prasad, N., Namdeti, R., Baburao, G., et al., 2024. Central composite design for the removal of copper by an Adansonia digitate. Desalination and Water Treatment. 317, 100164.
- <span id="page-13-6"></span>[29] Namdeti, R., Joaquin, A.A., Al Amri, A.M.A.H.M., et al., 2022. Application of artificial neural networks and response surface methodology for dye removal by a novel biosorbent. Desalination and Water Treatment. 278, 263–272.